

Full Length Research Paper

# Land Degradation Monitoring in the Nile Delta of Egypt, using Remote Sensing and GIS

Shalaby, A<sup>1</sup>., R. R. Ali<sup>2</sup> and A.Gad<sup>3</sup>

1. *Researcher, head of land use department, NARSS, National Authority for Remote Sensing and Space Sciences (NARSS), 23 Joseph Tito Street, El-Nozha El-Gedida, P.O. Box : 1564 Alf Maskan, Cairo, Egypt.*
2. *Assistant Professor National Research Centre (NRC), El-Bohouth Street, Dokki, P.O.Box 12622, Cairo, Egypt*
3. *Professor, Head of Environmental Studies and Land Use Division, NARSS, Address: National Authority for Remote Sensing and Space Sciences (NARSS), 23 Joseph Tito Street, El-Nozha El-Gedida, P.O. Box : 1564 Alf Maskan, Cairo, Egypt*

*\*Corresponding Author: Shalaby Adel*

## ABSTRACT

Land degradation is one of the main problems that threaten the limited highly fertile land in the Nile Delta of Egypt. In this research, Landsat TM satellite images of 1984 and 1992, and ETM<sup>+</sup> of 2006 have been used to study the impact of urban sprawl on agricultural land in the Nile Delta. Visual interpretation using on screen digitizing and change detection techniques were applied for monitoring the urban sprawl. 42 soil profiles have been collected, morphologically described and analyzed for the study of their chemical and physical properties. The rate of salinization, and water logging were estimated following the FAO methodology for assessing land degradation. As urban sprawl is a significant degradation process, multi-temporal satellite images were used for its detection. The annual rate of salt accumulation in the soils is 0.10 dS/m, where an area of 3103.2 km<sup>2</sup> (or 10.8 %) has a high salinity risk. The water table is decreased by a rate of 1.85 cm per year in the *Typic Aquisalids* soils and 0.37 cm per year in the *Typic Calcigypsid* soils. The water logged areas cover 7461.3 Km<sup>2</sup> (or 25.9 %) concentrated in the Northern part of the Delta. The results show that a total expansion of urban area was 202.6 Km<sup>2</sup> between 1992 and 2006 of which 37.41 km<sup>2</sup> were converted from the alluvial fertile soils and 165.2 km<sup>2</sup> from cultivated desert fringes. It can be concluded that salinization and the urban sprawl are the dominant degradation process in the Nile Delta.

**Keywords:** land degradation, urbanization, GIS, soil mapping, Nile delta, Egypt.

## INTRODUCTION

Land degradation is the process which lowers the current and / or the potential capability of soil to produce goods or services (FAO/UNEP, 1978). It can also be defined as "the decline in soil quality caused through its misuse by human" (Lal and Stewart, 1990) or "a process that describes human induced phenomena, which reduce the current and/ or future soil capability" (Ayoub, 1991). It implies long term decline in soil productivity and its environment-moderating capacity (Lal, 2001). The concept of soil degradation according to the Global Assessment of Soil Degradation (GLASOD) focuses on a human-induced process (UNEP, 1992).

Dryland environments comprise about 6.1 billion ha, or about 47% of all land area. Approximately 84% of this area falls within the arid, semiarid, and dry sub-humid climates which are inherently susceptible to desertification (UNEP, 1997). Estimates vary widely, but about 10-20% of the susceptible drylands is believed to have already undergone land degradation (Millenium Ecosystem Assessment, 2005). (Dudal, 1982; Rozanov, 1994) Estimated that about 70% of potentially productive dry lands, are currently threatened by various forms of land degradation.

Soil degradation is a severe global problem of modern times (Lal, 1998). (Steiner, 1996) noted that every year 5 to 7 million hectares of agricultural land worldwide become unproductive due to physical and chemical degradation. The problem is much more serious in arid and semi-arid regions than in tropical and temperate areas since arid soils are more prone to degradation because of their properties and the prevalent climatic conditions. The World Bank (1992, p. 57), states that salinization caused by inappropriate irrigation practices affects about 60 million ha, or 24% of all irrigated land. Severe declines in productivity are observed on about 24 million ha, or 10% of irrigated land.

(Coote *et al.*, 1982) reported that the processes which contribute to land degradation can be differentiated in two broad groups, those which can be observed over a wide area and are the result of regional soil management and climatic factors, and those which are localized in extent and are functions of local soil management conditions. Many of these processes are natural and some would proceed even without man's intervention. The water logging and subsequent salinization and/or alkalinization are the major land

degradation processes in irrigated agriculture lands of arid and semi-arid conditions (Dwivedi *et al.*, 1999). These processes can often be accelerated by the shifting of natural equilibrium, or by the inhabitation of natural soil-building processes. Two categories of soil degradation were recognized by (Oldeman, 1994). The first category is soil degradation due to displacement of soil material such as soil erosion by wind and water. The second category is in situ soil degradation due to chemical processes like loss of nutrients and organic matter, salinization, acidification, and pollution and due to physical processes such as compaction, waterlogging and subsidence.

Egypt's Nile Valley and Delta, one of the oldest agricultural areas in the world, has been under continuous cultivation for at least 7000 years. The total cultivated area of the Nile Delta is 1,828,840 hectares, representing 56.49% of the cultivated area of Egypt. The deltaic lands, worldwide, could be considered as problematic lands regarding their fine texture, poor drainage, as well as the intrusion of sea water to the ground water table. Therefore, these soils are threatened by land degradation processes more than any other alluvial soils. Land degradation processes are clearly observed also in different regions under the arid and semi-arid climatic conditions. The Nile Delta is one of these regions which are threatened by water logging, soil compaction, salinization and alkalinization.

In Egypt, construction of the Aswan High Dam in the 1960s enabled farmers in the Nile Valley and Delta to convert from basin to perennial irrigation, causing a rapid rise in shallow water tables and increasing soil salinity (Abul-Ata, 1977; Kinawy, 1977; Kishk, 1986; White, 1988; Scott, 1993). Within 20 years of the start of Aswan operations, waterlogging and salinization had affected 28% of farmland in Egypt and average yields in those areas had fallen by 30% (Speece and Wilkinson, 1982).

Urbanization is an inevitable process due economic development and rapid population growth. Encroachment of urban settlements on agricultural lands may pose dire consequences. The continuous increase of population density causes an increasing pressure on areas already inhabited and caused a decrease in area per capita from 0.12 ha in 1950 to 0.06 ha in 1990 (Suliman, 1991) and to 0.04 ha in 2009 (CAPMAS, 2009). Therefore, determining the trend and the rate of land cover conversion are necessary for the development planner in order to establish rational land use policy (Shalaby and Tateishi, 2007).

Remote sensing has long been used to study and map land degradation. (Lantieri, 2006) indicated that salinization patterns in irrigation schemes can be detected as salt appears as white patches. It is possible to detect bare surfaces and land use changes (through multi-temporal analysis) that may help to assess land degradation problems in an area. When vegetation cover maps are created for different years, it is possible to determine the changes in land cover, which may indicate vegetation degradation (Shalaby and Tateishi, 2007). This straightforward method of using vegetation cover maps

to assess desertification is robust and rather accurate, but remains restricted to a physical state assessment of desertification. It does not look at driving forces, and therefore it identifies the problem, may assess its intensity, but does not provide guidance to its solution (Lantieri, 2006).

Usually studies of land degradation combine satellite remote sensing information with other spatial data, like topography, soils, and land use, into a GIS. A GIS enables analysis of combinations of different data layers, which may result in a better understanding of land degradation problems, causes and consequences. Also, within a GIS modeling with relatively simple empirical and semi-empirical models can be done to determine the risk of land degradation in an area. For instance, (Okoth, 2003) used a simple logit regression equation that combines the parameters slope and ground cover to determine water erosion risk in Kiambu District of Kenya. The ground cover and slope data were derived from remote sensing imagery, while the erosion data for the regression analysis came from field studies. In the GIS, the regression model was used to determine the erosion risk in the entire area. A similar approach was used by (Vrieling *et al.*, 2002) to determine water erosion risk in the Colombian Eastern Plains. Instead of a simple erosion model, they used a decision tree for erosion classification. The decision tree combined relevant information on soils, slopes and vegetation cover, and was derived from expert knowledge

The main objective of this study is to evaluate the hazard of salinization, water logging and urban sprawl over the Nile Delta using remote sensing and soils database.

## MATERIALS

**Soil maps:** The soil maps of Egypt (ASRT, 1982) are the main materials collected and converted to the digital format. The collected soil maps cover the alluvial arable land and their interference with Desert fringes. These maps include a number of 12 map sheets at scale of 1: 100,000 in analogue format, covering the whole of Nile Delta region.

**Topographic maps:** Topographic maps at scale 1: 50,000 covering the Nile Delta, produced by the Egyptian General Survey Authority (EGSA) were converted to a digital format.

**Satellite data:** A number of 21 Landsat satellite images were used in this study to cover the study area in three different dates (1984, 1992 and 2006) to study the urban sprawl in the Nile delta.

## METHODS

### Geometric correction

Accurate per-pixel registration of multi-temporal satellite data is essential for change detection because registration errors could be inferred as land use/cover changes, leading to an overestimation of actual change (Stow, 1999). Change detection analysis is performed on a pixel-by-pixel basis;

therefore any misregistration greater than one pixel will provide an anomalous result for that pixel. To overcome this problem, the root mean square error (RMSE) between any two dates should not exceed 0.5 pixel (Lunetta and Elvidge, 1998). In this study, geometric correction was carried out using ground control points from digital topographic maps (UTM, WGS84) to geocode the image of 1992, then this image was used to register all the other images; the RMSE between different images was less than 0.4 pixel, which is acceptable.

### Image enhancement and visual interpretation

The goal of image enhancement is to improve the visual interpretability of an image by increasing the apparent distinction between the features. The process of visually interpreting digitally enhanced imagery attempts to optimize the complementary abilities of the human mind and the computer. The mind is excellent at interpreting spatial attributes on an image and is capable of identifying obscure or subtle features (Lillesand and Kiefer, 1994). Contrast stretching was applied on all images and the false colour composites (FCC) were produced. These FCC are visually interpreted using on-screen digitizing to delineate urban areas in the three different dates.

### Digital soil mapping

The soil map of the study area was extracted from the available soil map of Egypt produced by the Academy of Scientific Research and Technology (ASRT 1982); the original nomenclature of soil order, suborders and great groups has been updated according to the latest American Soil Taxonomy of USDA (2010). The transformation of the soil map (produced in 1982) into a digital format was done. The study area is covered by 12 soil map sheets. These sheets were scanned and geometrically corrected using UTM coordinate system and WGS84 projection. On-screen digitizing was used to convert the 12 sheets into vector formats and then edge matching was performed using ArcGIS 9.2. It was noticed, after edge matching, that there was a deviation (constant in many places in its direction and magnitude) between the produced map and the well-registered landmarks, driven out of

the survey maps and satellite images of the study area. After investigating this deviation, it has been attributed to two reasons: lack of a coordinate and the rubber-sheeting that accompanied the edge-matching task (Boullie 1978). To overcome this problem another spatial adjustment (transformation) has been performed. Well-registered topographic maps and accurately geo-referenced satellite images have been used to perform the transformation process. The transformation tools of ArcGIS 9.2 system were found to be effective in performing the spatial adjustment of the thematic map (Burrough 1986). A semi-detailed survey was done throughout the investigated area to gain an appreciation on the soil patterns, the land forms and landscape characteristics. The laboratory analysis of the study area reported by ASRT (1982) has been compiled in the database and incorporated into the attribute table of the soil map (Pavasovic 1993; Nguyen 2001). Urban land cover class was overlaid on the soil map, and then the areas lost from different soil due to urban sprawl were calculated.

### Field work and laboratory analyses

Field missions were carried out to represent the different soil units. A number of 42 soil profiles were studied in the field according to FAO guidelines, (FAO, 2006). Soil color was defined by Munsell Color Charts (Soil Survey Staff, 1975). The soil laboratory analyses (i.e. soil texture, CaCO<sub>3</sub> content, CEC, EC, ESP, pH, soluble cations and anions and organic matter content) were carried out for 95 disturbed soil samples, representing different soil profile horizons, using the soil survey laboratory methods manual (USDA, 2004). The results of these analyses have been compiled in the database and then incorporated into the attribute tables of the digital GIS soil maps (Nguyen Quec Dinh, 2001 and Pavasovic, 1993).

### Land degradation hazard assessment

Salinity and water logging hazard over the Nile Delta region was assessed using FAO/UNEP (1979) methodology for assessing soil degradation. The results were evaluated and confirmed with the soil units. The ratings used are present in Tables (1&2).

**Table (1)** Soil degradation classes and rates.

<b>Chemical degradation</b>	<b>Salinization (Cs) increase in (EC) per dS/m/year</b>
Non to slight	<0.5
Moderate	0.5-3
High	3-5
Very high	>5
<b>Physical degradation</b>	<b>Water logging/increase in water table in cm/year</b>
Non to slight	<1
Moderate	1-3
High	3-5
Very high	>5

Adapted from FAO (1979)

**Table (2)** Criteria used to determine the degree of the land degradation.

Hazard type	Indicator	Unit	Hazard class			
			Low	Moderate	High	Very high
Salinization	EC	dS/m	4	4-8	8-16	>16
Water Logging	Water table	Cm	150	150-100	100-50	<50

### Urban sprawl detection

Regardless of the technique used, the success of change detection from imagery will depend on both the nature of the change involved and the success of the image preprocessing and the classification procedures. If the nature of the change within a particular scene is either abrupt or at a scale appropriate to the imagery collected then change should be relatively easy to detect; problems occur only if spatial change is subtly distributed and hence not obvious within any image pixel (Milne 1988). In the case of the study area chosen, field observation and measurements have showed that the change in land cover between the three dates was both marked and abrupt. In this study post-classification change detection technique was applied. Post-classification is the most obvious method of change detection, which requires the comparison of independently produced classified images. Post-classification comparison proved to be the most effective technique, because data from two dates are separately classified, thereby minimizing the problem of normalizing for atmospheric and sensor differences between different dates. Urban land cover class was extracted from the visual interpretation of the satellite images, and then cross-tabulation analysis was carried out to study the spatial distribution and areas of urban sprawl (of 1984, 1992 and 2006) on different soil types, ArcGIS 9.2 software was used for this function.

## RESULTS AND DISCUSSION

### Soils of the Nile Delta

The soil map of the Nile Delta was extracted from the soil map of Egypt (ASRT, 1982), which was classified using the American Soil Taxonomy (USDA, 1975). The produced map (Fig. 1) has been updated according to the latest edition

(USDA, 2010). The obtained results indicate that the *Vertic Torrifluvents* is the major sub great group in the alluvial Nile Delta, covering 12,940.3 km<sup>2</sup>, representing 41.2% of the total area. Patches of sub-great groups *Typic Torrifluvents* are included in the alluvial soils. The sub-great group *Typic Quartzipsamments* dominates the eastern and western desert fringes, covering an area of 4,082 km<sup>2</sup> (12.9% of total area) in addition to an area of 375.2 km<sup>2</sup> interfered with the *Typic Torriorthents* sub-great group.

The *Typic Torripsamments/Typic Torriorthents* sub-great groups association exhibits an area of 399.1 km<sup>2</sup>, representing 1.3% of total area, adjacent to the previous one. The sub-great groups *Typic Torrifluvents*, *Typic Torriorthents* and *Typic Torripsamments* exhibit significant coverage in delta borders, as their areas are 2,514.9, 2,390.9 and 1,385.0 km<sup>2</sup> representing 8%, 7.6% and 4.4% of total area respectively. The sub-great groups *Typic Aquisalids*, *Typic Haplosalids* and *Aquic Torrifluvents* exist in the northern Delta region influenced by the northern Delta lakes. They cover areas of 1,928.8, 166.9 and 437.9 km<sup>2</sup>, representing 6.1%, 0.5% and 1.4% of the total Nile Delta area respectively.

Minor areas of *Typic Haplocalcids*, *Typic Haplogypsids*, *Typic Petrogypsids* and *Typic Calcigypsids* are distributed in the eastern and western Delta regions covering areas of 1,012.6, 1,007.5, 145.9 and 4.1 km<sup>2</sup> representing 3.2%, 3.2%, 0.5% and 0.01% respectively. Limited Hill and Rockland areas exist in the southern Delta region, covering areas of 1,223.2 and 1,426.9 km<sup>2</sup> respectively. Table (3) illustrates areas and representative soil profiles of the different soils in the mapped area.

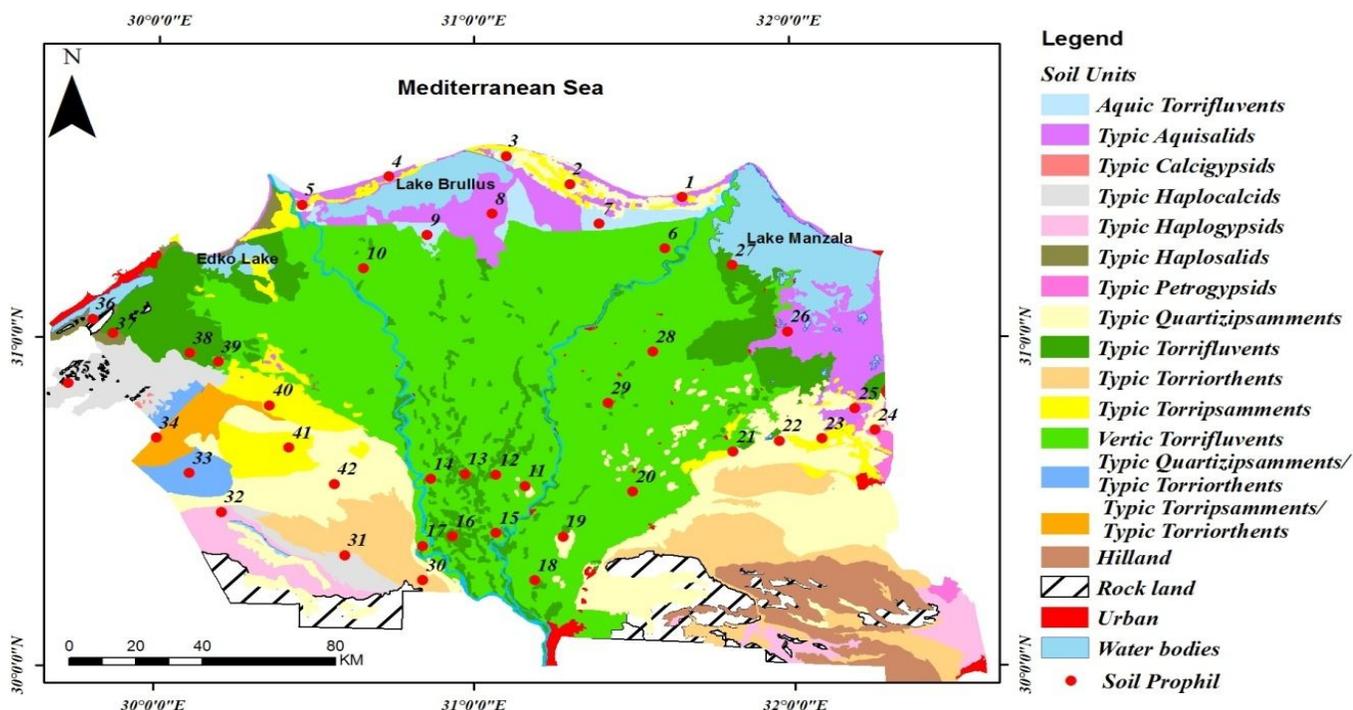


Figure (1) Soils of the Nile Delta overlaid by the representative soil profiles

Table (3) Areas, frequency and representative soil profiles of Sub great groups of the Nile Delta region

Type (Sub great group)	Area (Km <sup>2</sup> )	%	Representative soil profiles
<i>Aquic Torrifuvents</i>	437.94	1.39	<b>7</b>
<i>Typic Aquisalids</i>	1928.75	6.13	1, 4, <b>8</b> , 26
<i>Typic Calcigypsids</i>	4.09	0.01	<b>35</b>
<i>Typic Haplocalcids</i>	1012.60	3.22	<b>31</b> , 32
<i>Typic Haplogypsids</i>	1007.53	3.20	<b>25</b>
<i>Typic Haplosalids</i>	166.91	0.53	3, 9, <b>37</b>
<i>Typic Petrogypsids</i>	145.85	0.46	<b>24</b>
<i>Typic Quartzipsamments</i>	4082.00	12.98	2, 5, 11, 19, 22 and <b>42</b>
<i>Typic Quartzipsamments/ Typic Torriorthents</i>	375.18	1.19	<b>33</b>
<i>Typic Torripsamments/ Typic Torriorthents</i>	399.06	1.27	<b>34</b>
<i>Typic Torrifuvents</i>	2514.98	8.00	16, <b>13</b> , 18, 29, 27 and 38
<i>Typic Torriorthents</i>	2390.99	7.60	21, <b>30</b> ,
<i>Typic Torripsamments</i>	1385.04	4.41	23, 40 and <b>41</b>
<i>Vertic Torrifuvents</i>	12940.34	41.16	6, <b>10</b> , 12,14, 15, 17, 20, 28 and 39
Hilland	1223.24	3.89	--
Rock land	1426.88	4.54	36 (observation point)
Total	28791.26	100.00	

Note: Bold numbers are the typical soil profiles considered in assessing soil degradation rate and hazard

**Degradation rate and risk assessment**

Thirteen soil profiles were selected out of a total 42 studied profiles representing different soils of the Nile Delta. Soil

laboratory analyses (Table 4) were used to assess land degradation rate and hazard for each soil sub-group.

**Table (4)** Laboratory analyses of the typical representative soil profiles

Taxonomic unit	Representative profile	Depth in cm	Texture class	pH 1:02.5	EC dS/m	CaC O <sub>3</sub> %	OM %	ESP	CEC meq/100 g. soil
<i>Aquic Torrifluvents</i>	7	0 - 10	C	7.73	3.68	2.51	1.39	12.5	46.5
		25-10	C	7.81	2.83	2.31	1.7	11.7	48.8
		25 - 40	CL	7.88	2.56	1.62	0.31	14.3	40.3
		40 - 60	CL	7.91	1.78	1.12	0.38	14.3	39.4
<i>Typic Aquisalids</i>	8	0-25	C	7.81	13.3	1.33	1.7	13.2	52.6
		25-50	C	7.88	11.04	2.05	0.3	12.6	43.2
<i>Typic Haplosalids</i>	37	0 - 30	SL	7.95	11.7	5.55	0.39	16.2	14.5
		30 - 50	SL	7.87	18.9	4.41	0.64	14.8	17.4
		50 - 80	SCL	8.05	30.6	7.22	0.41	15.6	32.4
		80 - 110	CL	7.64	16.2	9.71	0.42	13.4	38.2
<i>Typic Torrifluvents</i>	13	0-20	CL	8.15	3.2	3.15	2.86	16.5	40.6
		20-45	CL	8.4	1.32	2.11	1.35	13.7	39.3
		45-80	C	8.01	1.8	1.6	1.91	14.5	44.3
		80-110	C	8.13	3.57	1.2	2.12	12.8	45.8
<i>Vertic Torrifluvents</i>	10	0-10	SiC	8.4	0.92	2.11	2.01	12.3	42.3
		10-35	SiC	8.5	0.79	1.03	0.7	14.5	47.6
		35-70	SiC	8.25	0.62	0.92	1.42	13.2	49.8
		70-120	C	8.34	0.8	0.88	1.77	15.8	45.1
<i>Typic Calcigypsids</i>	35	0 - 15	C	7.57	10.9	7.52	0.51	11.3	45.6
		15 - 40	CL	7.74	6.1	8.61	0.57	14.7	37.4
		40 - 60	SCL	7.7	5.8	6.92	0.45	14.9	27.8
		60 - 90	LS	7.86	2.52	9.52	0.23	15.7	14.8
<i>Typic Haplocalcids</i>	31	0-25	SL	7.92	3.65	13.52	0.12	25.6	12.8
		25-70	LS	8.1	4.22	14.23	0.15	26.4	8.6
<i>Typic Haplogypsids</i>	25	0 - 15	SL	8.21	14.23	10.22	1.65	22.6	19.6
		15 - 50	SCL	8.41	11.65	8.65	1.11	24.2	30.5
		50 - 85	SCL	7.96	14.71	7.33	0.42	26.4	28.6
<i>Typic Petrogypsids</i>	24	0-20	CL	8.21	7.51	9.12	0.51	16.84	28.3
		20-75	CL	8.05	8.11	11.23	0.12	21.61	46.1
<i>Typic Torriorthents</i>	30	0-15	L	8.23	3.15	4.86	1.13	13.7	23.6
		15-35	L	8.41	4.33	5.19	1.52	11.9	25.4
		35-75	C	8.11	5.62	6.73	0.41	12.7	45.3
		75-130	C	8.05	7.11	7.12	0.1	18.6	39.5
<i>Typic Torripsammets</i>	41	0 - 15	S	7.6	7.62	1.51	0.41	14.8	8.9
		15 - 40	S	7.93	3.49	1.95	0.21	12.6	9.5
		40 - 130	S	8	3.51	2.03	0.02	13.5	9.3
<i>Typic Quartzipsammets</i>	42	0-25	S	8.1	0.77	8.33	0.3	11.2	7.2
		25-100	S	8.57	0.2	10.23	0.32	10.3	8.3
<i>Typic Quartzipsammets/Torriorthents</i>	33	0 - 15	LS	7.72	5.03	7.74	0.36	12.5	12.3
		15 - 50	S	7.73	7.08	9.15	0.32	14.3	10.6
		50 - 120	S	7.61	5.92	3.65	0.05	9.8	11.4

The rate of dominant land degradation processes (i.e. water logging and salinity) was estimated by the comparison between the main land characteristics of (1982) and (2009). Consequently the rates of land degradation, for each type, were classified to slight, moderate or high. The risk of water logging and salinity was estimated from the present value of electric conductivity, and the depth of water table. The rate and risk of water logging and salinity of the Nile Delta soils are presented in table (5). The obtained data were attached to the attribute table of the digital soil map of Nile Delta for mapping the water logging and salinity hazard. Figures (2 & 3) and table (6) represent the spatial distribution and areas of water logging and salinity hazard classes over the Nile Delta region. The obtained data reveals the following:

#### Water logging:

The results show that the soils depth of the studied area varies from 50 cm to 130 cm. The soils of *Aquic Torrifluvents*, *Typic Aquisalids*, *Typic Calcigypsids*, *Typic Haplocalcids*, *Typic Haplogypsids*, *Typic Petrogypsids*, are characterized by a soil depth differ from 50 to 90 cm, these soils are threatened by a moderate to high water logging risk. The rest of the soils have

a depth ranging from 100 to 130 cm, accordingly characterized by slight risk of water logging. The annual decrease of water table reach its maximum in the *Typic Aquisalids* and *Typic Quartzipsamments* (1.85 cm/year), while recording 1.48 cm/year in the *Aquic Torrifluvents* and 1.11 cm/year in *Vertic Torrifluvents*, *Typic Haplocalcids* and *Typic Quartzipsamments/ Typic Torriorthents* soils. The soils of moderate rate and very high risk of water logging (M, 4) occupy an area of 1,928.7 km<sup>2</sup> (i. e. 6.70 %) represented by *Typic Aquisalids* soils in the north of the Nile Delta. The soils of *Aquic Torrifluvents*, *Typic Haplocalcids* and *Typic Quartzipsamments* are characterized by moderate rate and high risk of water logging (M, 3) covering an area of 5,532.5 km<sup>2</sup> (i. e. 19.2 %). These soils are found mainly at the western and eastern parts of the Delta, also a small strip is found at the north of the Nile Delta near to El Borlolus Lake. Most of the alluvial deposits of the Delta are characterized by moderate rate and risk of water logging associating the *Vertic Torrifluvents*, *Typic Quartzipsamments/Typic Torriorthents* and *Typic Torripsamments/ Typic Torriorthents* soils, representing 47.6 % of the total area.

**Table (5)** Water logging and salinity (rate & risk) of the different soils

Taxonomic unit	Representative profiles	Water logging (depth in cm)					Salinity average (dS/m)				
		1982	2009	Rate	Class	Risk	1982	2009	Rate	Class	Risk
<i>Aquic Torrifluvents</i>	7	100	60	1.48	M	3	0.8	2.71	0.07	S	1
<i>Typic Aquisalids</i>	8	100	50	1.85	M	4	8.5	12.17	0.14	S	3
<i>Typic Haplocalcids</i>	37	130	110	0.74	S	2	10.5	19.35	0.33	S	4
<i>Typic Torrifluvents</i>	13	130	110	0.74	S	2	1.4	2.47	0.04	S	1
<i>Vertic Torrifluvents</i>	10	150	120	1.11	M	2	0.5	0.78	0.01	S	1
<i>Typic Calcigypsids</i>	35	100	90	0.37	S	3	2.3	6.33	0.15	S	2
<i>Typic Haplocalcids</i>	31	100	70	1.11	M	3	1.5	3.93	0.09	S	1
<i>Typic Haplogypsids</i>	25	100	85	0.56	S	3	7.6	13.53	0.22	S	3
<i>Typic Petrogypsids</i>	24	100	75	0.93	S	3	4.6	7.81	0.12	S	2
<i>Typic Torriorthents</i>	30	150	130	0.74	S	2	4.2	5.05	0.03	S	2
<i>Typic Torripsamments</i>	41	150	130	0.74	S	2	1.6	4.87	0.12	S	2
<i>Typic Quartzipsamments</i>	42	150	100	1.85	M	3	1.2	0.49	0.00	I	1

<i>Typic Quartzipsamments/Typic Torriorthents</i>	33	150	120	1.11	M	2	4.5	6.01	0.05	S	2
<i>Typic Torripsamments/Typic Torriorthents</i>	34	150	110	1.48	M	2	15.1	15.23	0.00	N	3

Note: the data of 1982 was extracted from "soil map of Egypt" after ASRT, 1982

Degradation Rate: I= improved, N= non, S= slight, M=moderate

Degradation risk: 1= low, 2= moderate, 3= high, 4= very high

The slight rates of water logging are found associated with the *Typic Haplosalids*, *Typic Torrifluvents*, *Typic Torriorthents*, and *Typic Torripsamments* soils, they also have a moderate risk (S, 2) of water logging, exhibiting an area of 6,457.9 km<sup>2</sup> (i. e. 22.4 %). On the other hand, the soils of *Typic Haplogypsids* and *Typic Petrogypsids* have a slight rate of water logging but with a high risk (S, 3) found only in the eastern and western desert fringes of the Nile Delta, covering an area of 1,157.5 km<sup>2</sup> (i. e. 4.0 %).

**Soil Salinity:**

The values of electric conductivity (EC) vary widely in the study area, where the EC values range between 0.46 – 19.35 dS/m. that indicates variable levels of salinity risk. On the other hand, the rate of EC increase in different soils is generally slight (i.e. less than 0.5 dS/m per year), attributed to the existence of network of effective drains. This indicates that the salinity hazard in the Nile Delta is related mainly to its

current state. The very high risk of salinization (S, 4) is found in the northern parts of Nile Delta, associated with the marine and lacustrine deposits. This class occupies the *Typic Haplosalids* soils covering only 166.9 km<sup>2</sup> (i. e. 0.6 %). The soils of high salinity risk (S, 3) cover an area of 2,936.3 km<sup>2</sup> (i. e. 10.2 %) found associated with *Typic Aquisalids*, *Typic Haplogypsids* and *Typic Torripsamments/Typic Torriorthents* soils. The soils of *Typic Calcigypsids*, *Typic Petrogypsids*, *Typic Torriorthents*, and *Typic Torripsamments* are characterized by moderate risk of salinity. They cover an area of 4,301.1 km<sup>2</sup> (i. e 14.9 %) situated in the eastern and western desert fringes and the lacustrine deposits at the north of the Nile Delta. The soils of low risk dominate the whole alluvial deposits in the Nile Delta, covering an area of 16,905.9 km<sup>2</sup> (i.e. 58.7%). These classes of salinity risk are found associated with the *Aquic Torrifluvents*, *Typic Torrifluvents*, *Vertic Torrifluvents*, *Typic Haplocalcids*, and *Typic Quartzipsamments* soils.

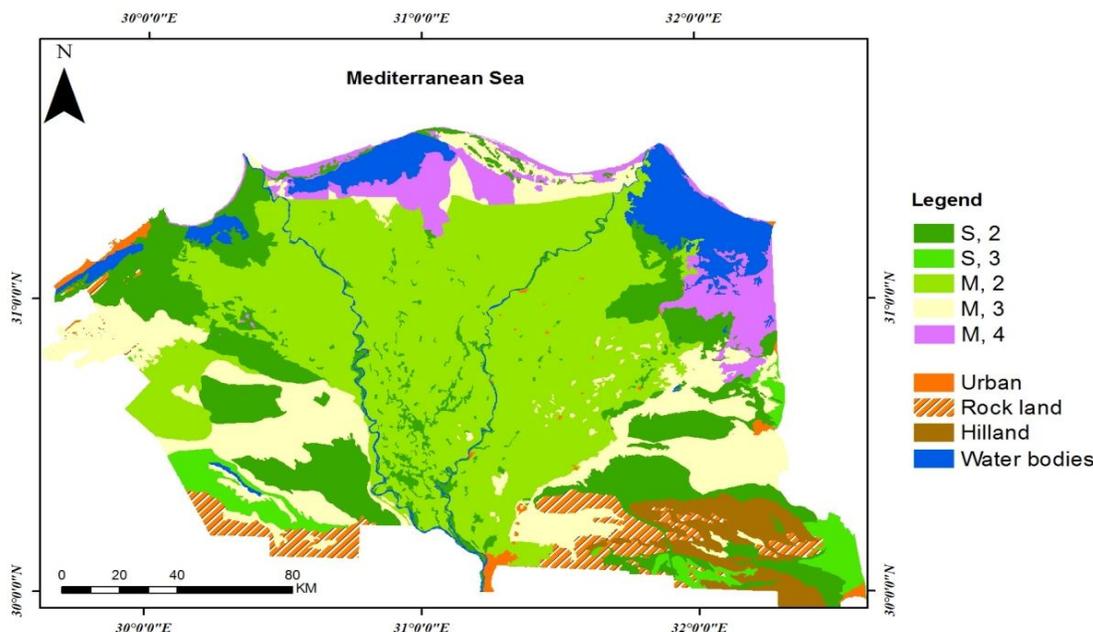


Figure (2) Water logging rate and risk over the Nile Delta

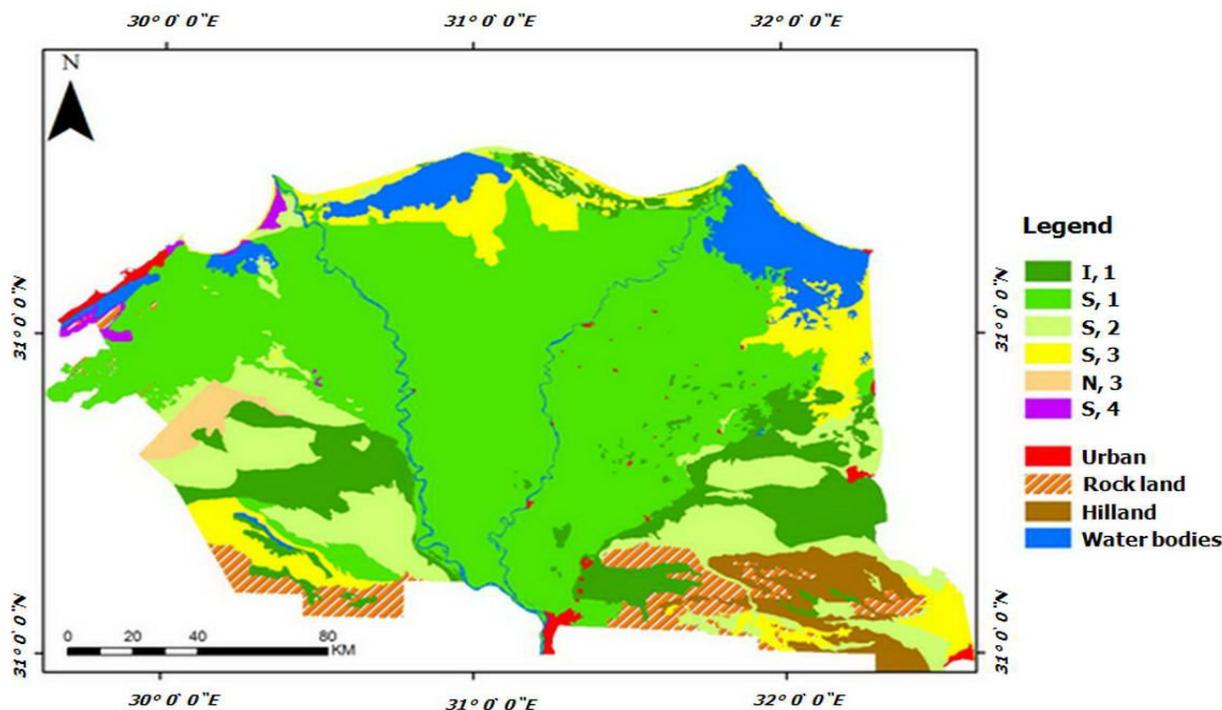


Figure (3) Salinity rate and risk over the Nile Delta

Table (6) Areas of the different classes of water logging and salinity over the Nile Delta

Water logging classes	Area (km <sup>2</sup> )	Area (%)	Salinity classes	Area (km <sup>2</sup> )	Area (%)
M, 2	13714.62	47.63	I, 1	4082.01	14.18
M, 3	5532.54	19.22	N, 3	399.06	1.39
M, 4	1928.73	6.70	S, 1	16905.89	58.72
S, 2	6457.90	22.43	S, 2	4301.13	14.94
S, 3	1157.47	4.02	S, 3	2936.26	10.20
--	--	--	S, 4	166.91	0.58
Total	28791.26	100.00	Total	28791.26	100.00

#### Urban sprawl:

Urban settlement over the Nile Delta, in 1984, 1992 and 2006, was mapped and change detection technique resulted in following the evolution of urban sprawl (Figure 4). The impact of this urban sprawl on agricultural land was evaluated and the statistical data, representing the spatial urban changes from 1984 to 2006 are illustrated in Table 7. The built-up

areas in the Nile Delta increased from 1134.7 km<sup>2</sup> in the year 1984 to 1593.7 km<sup>2</sup> in 1992 and to 3671.0 km<sup>2</sup> in the year 2006. The obtained data indicate that the urban expansion during the 1984 – 2006 was on the expense of the most fertile soils where, the lost *Vertic Torrifuvents* area is 797.9 km<sup>2</sup>, *Typic Torrifuvents* lost 307.6 km<sup>2</sup> and the *Typic Torriorhents* soils lost an area of 320.8 km<sup>2</sup>.

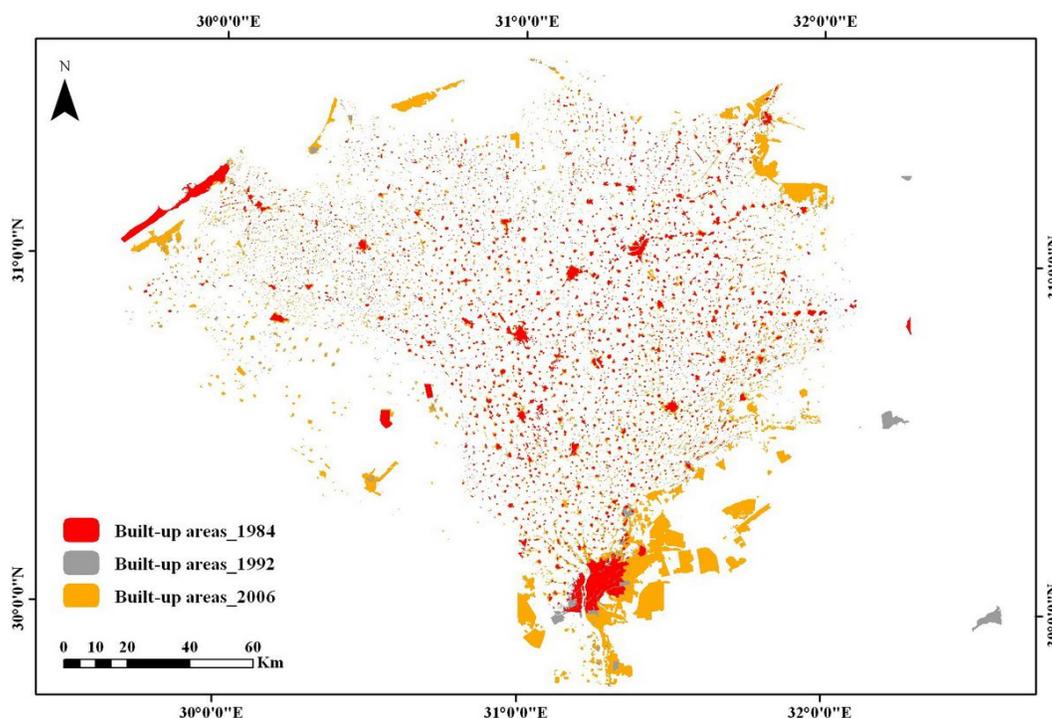


Figure (4) Urban settlements in the Nile Delta in 1984, 1992 and 2006

Table (7) Changes in the areas of different soils and urban in the Nile Delta region in 1984, 1992 and 2006

Mapping unit	1984	1992	2006	A	B	C
<i>Aquic Torrifluvents</i>	426.76	421.63	402.77	-5.14	-18.85	-23.99
Hilland	441.00	441.00	427.99	-0.00	-13.01	-13.01
Rock land	2315.20	2291.49	1931.24	-23.71	-360.24	-383.95
<i>Typic Aquisalids</i>	1921.48	1910.18	1806.93	-11.30	-103.26	-114.56
<i>Typic Calcigypsid</i>	4.09	3.99	3.61	-0.09	-0.38	-0.47
<i>Typic Haplocalcids</i>	1064.22	1055.80	945.71	-8.42	-110.09	-118.51
<i>Typic Haplogypsid</i>	1034.21	1007.53	997.58	-26.69	-9.94	-36.63
<i>Typic Haplosalids</i>	166.73	162.99	126.91	-3.74	-36.07	-39.82
<i>Typic Petrogypsid</i>	145.85	145.85	145.85	0.00	0.00	0.00
<i>Typic Quartizipsamments</i>	5051.77	5009.45	4786.59	-42.32	-222.86	-265.18
<i>Typic Quartizipsamments/ Typic Torriorthents</i>	374.93	374.36	366.52	-0.57	-7.83	-8.41
<i>Typic Torrifluvents</i>	1368.04	1128.16	1060.69	-239.88	-67.46	-307.35
<i>Typic Torriorthents</i>	2951.55	2933.25	2630.79	-18.30	-302.47	-320.76
<i>Typic Torripsamments</i>	809.81	775.31	749.63	-34.50	-25.68	-60.18
<i>Vertic Torrifluvents</i>	14122.12	14075.93	13324.17	-46.18	-751.76	-797.94
Urban	1132.85	1593.69	3623.59	460.83	2029.90	2490.74
Total	33330.60	33330.60	33330.60			

Note: A= the difference between 1984 and 1992, B= the difference between 1992 and 2006 and C= the difference between 1984 and 2006.

Most of the urban expansion over the fertile soils happened during the 1992 to 2006 period. From 1984 to 1992 the *Vertic Torrifuvents*, soils lost a relatively small areas (i.e. 46.2 km<sup>2</sup>) compared with 751.8 km<sup>2</sup> during 1992 – 2006 period. The areas of Rockland and hilland was stable during 1984 – 1992 period, where the lost area amounted only 23.7 km<sup>2</sup> compared with 383.9 km<sup>2</sup> during 1992 – 2006. The latest urban expansion in the desert fringes of the Nile Delta between 1992

## CONCLUSION

The objective of this study was to evaluate the hazard of salinization, water logging and urban sprawl over the Nile Delta using remote sensing and soils database. It can be concluded that the salinization and urban sprawl are the dominant land degradation processes in the Nile Delta. Application of GIS and analyzing multi-temporal satellite images are found as elaborative techniques for assessing the current status and land degradation hazard indices.

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